

HUMAN HEALTH RISK ASSESSMENT FORMER CELOTEX SITE 2800 S. SACRAMENTO AVE. CHICAGO, ILLINOIS

Prepared for City of Chicago Department of Environment 30 N. LaSalle Street - 25th Floor Chicago, Illinois 60602

February 7, 2007



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Subject: City of Chicago Sampling Data

Former Celotex Site – 2800 South Sacramento Avenue

To Whom It May Concern:

The City of Chicago Department of Environment is providing the following environmental reports for inclusion into the existing repository at the Marshall Square Branch Library, maintained by the U.S. Environmental Protection Agency, for the Former Celotex Site, located at 2800 South Sacramento Avenue:

- Former Celotex Site 2800 S. Sacramento Avenue, Phase II Subsurface Investigation, prepared by URS, dated November 29, 2006
- Human Health Risk Assessment, Former Celotex Site, 2800 S. Sacramento Ave., Chicago, Illinois, prepared by URS, dated February 7, 2007.

If you have questions or need additional information, please do not hesitate to contact me at (312) 744-3636.

Sincerely,

Leigh E. Peters, P.E.

Environmental Engineer III

cc: Jena Sleboda, USEPA





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Former Celotex Site - Human Health Risk Assessment

FIGURES

Figure 1 Sample Location Map

ATTACHMENTS

Attachment A Site Conceptual Exposure Model

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ACRONYMS

ATSDR Agency for Toxic Substances Disease Registry

CDOE Chicago Department of Environment

COC Chemical of Concern

cPAH Carcinogenic Polycyclic Aromatic Hydrocarbon

C_{sat} Saturation Concentration

ED Exposure Duration
EF Exposure Frequency

EFH Exposure Factors Handbook EPC Exposure Point Concentration

HHRA Human Health Risk Assessment HSDB Hazardous Substance Data Bank

IAC Illinois Administrative Code

IEPA Illinois Environmental Protection Agency

IRIS Integrated Risk Information System

Koc Organic Carbon Partition Coefficient
Kow Octanol Water Partition Coefficient

MOU Memorandum of Understanding

NCEA National Center for Environmental Assessment

PCB Polychlorinated Biphenyl PRG Preliminary Remediation Goal

RAGS Risk Assessment Guidance for Superfund

RfD Reference Dose RFS Request for Services RO Remediation Objective

SCEM Site Conceptual Exposure Model

SF Slope Factor

SSL Soil Screening Level

TACO Tiered Approach to Correction Action Objectives

UCL Upper Confidence Limit

USEPA United States Environmental Protection Agency

Former Celotex Site - Human Health Risk Assessment

VOC Volatile Organic Compound

URS Corporation (URS) was directed by the Chicago Department of Environment (CDOE) to proceed with a human health risk assessment (HHRA) pursuant to CDOE's request for services (RFS) dated May 30, 2006. This HHRA is based upon the data collected by URS at the former Celotex property located at 2800 S. Sacramento Avenue (the Site) in Chicago, Illinois. The results of URS' investigation were presented in a letter report, dated November 15, 2006, and submitted to CDOE (URS, 2006). The HHRA was performed in accordance with relevant guidance provided by the United States Environmental Protection Agency (USEPA) and the Illinois Environmental Protection Agency in the following documents:

- Risk Assessment Guidance for Superfund (RAGS) (USEPA, 1989);
- Calculating Upper Confidence Limits for Exposure Point Concentrations at Hazardous Waste Sites (USEPA, 2002a);
- Supplemental Guidance for Developing Soil Screening Levels for Superfund Sites (USEPA, 2002b);
- Exposure Factors Handbook (EFH) (USEPA, 1997);
- RAGS Part E, Supplemental Guidance for Dermal Risk Assessment (USEPA, 2004a); and
- Tiered Approach to Corrective Action Objectives (TACO) (Illinois EPA, 2001) Illinois Administrative Code [IAC] Title 35, Subtitle G, Chapter I, Subchapter f, Part 742 (TACO Guidance).

The intent of this risk evaluation is to determine whether a layer of on-site gravel fill material is suitable for incorporation into an engineered barrier across the Site. It is URS' understanding that the barrier will ultimately allow the Site to be used for recreational purposes. Based on the future recreational land use of the Site, this HHRA was completed using the most conservative exposure scenario (residential) to determine if recreational receptors will be exposed to unacceptable concentrations at the Site. The residential land use scenario was used because it is more conservative and the exposure parameter values to estimate risk are widely accepted by both USEPA and IEPA. However, a recreational land use scenario was also evaluated for comparative purposes to determine risks that would be observed for a more realistic future recreational receptor.

1.0 TACO RISK SCREENING APPROACH

Although the presence of many chemicals may be identified in the environmental samples collected during site investigative activities, the results of an HHRA are typically driven by a few chemicals and exposure pathways. To streamline the HHRA process and focus efforts on important issues, several methods have been developed by the regulatory agencies and the scientific community for the identification of chemicals and pathways that contribute significantly to the total risks posed by a site. A tiered, risk-based approach was used for the

selection of COPCs to be further evaluated in the formal HHRA for the Site. This approach is based on USEPA-developed methodology and follows standard HHRA procedures (USEPA, 1989; USEPA, 2002b). For this Site, IEPA's TACO approach was used to determine the list of chemicals that are most likely to drive risk at the Site. TACO is the IEPA's method for developing risk-based remediation objectives (ROs) for contaminated soil and groundwater, with consideration of Site conditions and identified land use and are calculated using USEPA guidance. ROs are designed to protect human health. ROs were used as readily-available risk-based concentrations to determine the list of chemicals that are most likely to drive risk for the Site. ROs were selected because they address residential and construction land use scenarios that are relevant to this Site.

TACO provides three options to develop ROs, of which selection depends on site-specific conditions and remediation goals:

- Exclusion of an exposure pathway;
- Use of area background concentrations; and
- A three-tiered approach for deriving ROs.

The HHRA represents a Tier 3 evaluation for the Site and consists of the following components:

Site-Specific Site Conceptual Exposure Model (SCEM). Attachment A presents the SCEM for the Site. The SCEM provides physical-chemical properties, and fate and transport characteristics of chemicals of concern (COCs) which were identified in the Tier 1 evaluation (URS, 2006). The SCEM also identifies potential sources, migration pathways, potential receptors, and exposure routes for COCs at the Site.

Tier 3 Formal Risk Assessment A Tier 3 formal risk assessment for carcinogenic polycyclic aromatic hydrocarbons (cPAHs) in gravel fines at the Site is presented in this report.

Exposure Route Exclusion. Section 3.0 of this report discusses justification for excluding certain exposure routes at the Site in accordance with TACO. Particularly, Section 3.2 presents a demonstration for excluding the soil migration to the uppermost aquifer route at the Site.

2.0 TIER 1 SCREENING RESULTS

A Tier 1 screening evaluation was conducted using laboratory data from gravel fines samples collected at the Site. The Tier 1 screening was performed based on the residential land use scenario, and the Tier 1 ROs were obtained from Appendix B, Table A of TACO (35 IAC 742). The soil ROs given in TACO are associated with three exposure routes: soil ingestion, soil inhalation, and soil migration-to-groundwater. All three exposure routes were used to develop the list of chemicals exceeding Tier 1 ROs.



Tier 1 screening indicates chemicals exceeding TACO Tier 1 soil ROs include methylene chloride, cPAHs (benz[a]anthracene, benzo[a]pyrene, benzo[b]fluoranthene, dibenz[a,h]anthracene, and indeno[1,2,3-cd]pyrene), dieldrin, and chromium. Methylene chloride, cPAHs, dieldrin and chromium exceed the Tier 1 soil migration-to-groundwater ROs. The cPAHs exceed the Tier 1 residential soil ROs. Table 1 presents a summary of the Tier 1 RO exceedances.

Approaches to Address Tier 1 Exceedances Identified at the Site

The following approaches will be taken to address the Tier 1 exceedances identified at the Site.

Area/Media	COCs	Tier 1 ROs Exceeded	Approach to Address Tier 1 RO Exceedances	Addressed in
Gravel Fines/Fill	Benzo(a)anthracene Benzo(a)pyrene Benzo(b)fluoranthene Dibenz(a,h)anthracene Indeno(1,2,3-cd)pyrene Methylene chloride Dieldrin Chromium	Soil Migration-to- Groundwater	Exposure Route Exclusion	Section 3.0 of this Report
Gravel Fines/Fill	Benzo(a)anthracene Benzo(a)pyrene Benzo(b)fluoranthene Dibenz(a,h)anthracene Indeno(1,2,3-cd)pyrene	Residential Ingestion	Tier 3 Formal Risk Assessment	Section 4.0 of this Report

3.0 EXPOSURE ROUTE EXCLUSION

This section evaluates potential for excluding exposure routes at the Site in accordance with Title 35 of the Illinois Administrative Code 35 (35 IAC 742), Subparts C and I.

3.1 Criteria for Exposure Route Exclusion

According to 35 IAC 742.300, the extent and concentrations of COCs must be characterized in order to evaluate pathways for exclusion. The extent and concentrations of COCs have been evaluated using data collected for this investigation (URS, 2006).

In addition, 35 IAC 742.305 specifies that no exposure route shall be excluded from consideration relative to a COC unless six additional requirements for demonstrating the absence of free product are met. These requirements are summarized in the table that follows.

35 IAC 742.305: Contaminant Source and Free Product Determination

Regulatory Requirement	Site Condition	Is Requirement Met?
(a) The sum of the concentrations of all organic COCs at each discrete sampling point shall not exceed the attenuation capacity of the soil (a default value of 6,000 mg/kg for soils within the top meter and 2,000 mg/kg for soils below one meter of the surface as set forth in 35 1AC 742.215);	The gravel fill sampling results indicated the soil attenuation capacity is not exceeded at the Site since the sum of the concentrations of all organic COCs at each sampling point did not exceed the lowest default natural organic carbon fraction (i.e., 6,000 mg/kg for surface, and 2,000 mg/kg for subsurface) as specified in 35 IAC 742.215 (b) (1) (A).	Yes
(b) The residual concentrations of any organic COCs remaining in the soil shall not exceed the soil saturation limit (C_{sat}) as determined under 35 IAC 742.220;	The concentrations of organic COCs were compared to the default C_{sat} values as given in Appendix A, Table A of TACO. No exceedances of C_{sat} were identified at the Site.	Yes
(c) Any soil which contains COCs shall not exhibit any of the characteristics of reactivity for hazardous waste as determined under 35 IAC 721.123;	No evidence exists to indicate that the fill materials exhibit any characteristics of reactivity as described in 35 IAC 721.	Yes
(d) Any soil which contains COCs shall not exhibit a pH less than or equal to 2.0 or greater than or equal to 12.5;	The pH levels of the gravel fill materials at the Site ranged from 8.9 to 11.9 (see Table 3 of URS, 2006).	Yes
(e) Any soil which contains COCs in the following list of inorganic chemicals or their salts shall not exhibit any of the characteristics of toxicity for hazardous waste as determined by 35 IAC 721.124, or an alternative method approved by the Agency: arsenic, barium, cadmium, chromium, lead, mercury, selenium or silver.	No unusually high total concentrations of inorganics were found, except for chromium levels detected in two samples taken from sample locations S-40 and S-41 at depths less than 2 feet. The detections of chromium are just slightly above the migration to groundwater RO. The average concentration (15.5 milligrams per kilogram [mg/kg]) and 95% upper confidence limit (UCL) (16 mg/kg) for chromium are below the migration-to-groundwater RO. UCL calculations are provided as Table B-1 in Attachment B.	Yes
(f) If COCs include polychlorinated biphenyls (PCBs), the concentration of any PCBs in the soil shall not exceed 50 parts per million as determined by SW846 Methods.	Fifty-nine samples were analyzed for PCBs. PCBs were not detected above 50 parts per million, residential or construction worker remedial objectives for soil ingestion in any of the 59 samples analyzed.	Yes

The conditions of the on-site gravel fill material meet all the criteria for exposure route exclusion set forth in 35 IAC 742.300 through 305. Exclusion of exposure routes can be considered an option at the Site.

3.2 Soil Migration-to-Groundwater Route Exclusion

This section provides a demonstration pursuant to 35 IAC 742.925 to exclude the soil migration-to-groundwater route for the Site. The evaluation was conducted in light of the following:

- Existence of an engineered barrier;
- City of Chicago Groundwater Ordinances limiting groundwater use; and
- Physical, chemical and migration properties of the COCs

35 IAC 742.925 Demonstration of Exposure Route Exclusion

As indicated in 35 IAC 742.300(c), TACO allows the exclusion of exposure routes under a Tier 3 evaluation as set forth in 35 IAC 742.925. The 35 IAC 742.925 outlines the items that need to be addressed under the Tier 3 evaluation to demonstrate that there is no actual or potential impact of COCs to receptors via a particular exposure route. Herein, it is demonstrated that there is no impact of COCs to receptors from the migration to groundwater route at the Site. The regulatory information and associated site conditions are outlined below.

Regulatory Requirement	Technical Demonstration
35 IAC 742.925 (a)	The route being evaluated is the soil migration-to-groundwater route.
A description of the route evaluated.	The discussion provided below in (b) through (d) demonstrates that potential groundwater receptors would not be impacted by COCs in fill material at the Site through the soil migration-to-groundwater route.
35 IAC 742.925 (b) Description of the site and physical site characteristics.	The Site comprises a rectangular-shaped parcel with smaller areas protruding from the central portion of the Site, and occupies approximately 18.26 acres. The Site is covered by a layer of gravel fill material at approximate depths of 0.67-1.7 feet below ground surface. The source of the gravel is not known. Soil was observed beneath the gravel fill during sampling activities. It is URS' understanding that these soils may also have been brought on-site from an unknown source. The Site is currently used for storage of trailers and other vehicles.
35 IAC 742.925 (c) Discussion of the result and possibility of the route becoming active in the future.	The City of Chicago provides a restriction on the use of groundwater. The ordinance as outlined in the Memorandum of Understanding Between the City of Chicago and the Illinois Environmental Protection Agency (IEPA, 1997) prohibits the installation of new potable water supply wells and that the potable water supply must be from an approved water distribution system. The provisions of this ordinance are applicable to the Site.
	Therefore, the soil migration-to-groundwater route will not become active in the future unless the City of Chicago groundwater ordinance is rescinded.
35 IAC 742.925 (d) (1) & (2)	This focus of the HHRA is on the gravel fines within the gravel surface cover at the Site.
Technical support including a discussion of the natural or manmade barriers to exposure through that route, calculations and modeling results.	It is URS' understanding that the current gravel surface will be used as an engineered barrier. It is also assumed that for public park construction, the gravel layer would be covered with soil and grass which effectively renders all pathways associated with exposure to gravel fines incomplete.
35 IAC 742.925 (d) (3) & (4) Physical and chemical and contaminant migration properties of contaminants of	Methylene chloride, cPAHs, dieldrin, and chromium were COCs at the Site that had soil migration-to-groundwater RO exceedances. The fate and transport characteristics of these COCs are described in the SCEM presented in Attachment A.
concern.	Significant leaching of cPAHs and dieldrin from on-site fill materials is not expected because of their low mobility and/or high sorption rates.
	Volatilization and biodegradation are the dominant transformation processes for methylene chloride. Significant impacts from methylene chloride in the on-site fill material to the deeper regional groundwater and the subsequent migration of methylene chloride in groundwater to reach the potential receptors are not likely because of biodegradation and volatilization. In

Regulatory Requirement	Technical Demonstration
	addition, as indicated in URS' letter report (URS, 2006), the presence of methylene chloride in the samples collected during the field investigation may be attributed to laboratory contaminants.

Conclusion Regarding Exposure Route Exclusion

The evaluation concluded no current or potential future receptors are impacted or will be impacted by the COCs in gravel fines at the Site through the soil migration-to-groundwater route. Therefore, the soil migration-to-groundwater route can be excluded from further evaluation as long as continued compliance with the groundwater ordinance is observed.

4.0 TIER 3 – FORMAL RISK ASSESSMENT FOR CARCINOGENIC PAHS

URS conducted a formal risk assessment to quantitatively evaluate the potential health impacts associated with exposure to cPAHs detected in gravel fines at the Site. Elevated levels above TACO Tier 1 soil ingestion ROs were detected in several samples collected from the Site for benz(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, dibenz(a,h)anthracene, and indeno(1,2,3-c,d)pyrene. However, all seven cPAHs were included in the risk assessment to address the additivity of these chemicals as they are considered similar-acting carcinogenic chemicals targeting the same organ (i.e., gastrointestinal system) according to 35 IAC 742 Appendix A, Table F.

The risk assessment was conducted in accordance with guidance provided in the following documents:

- Risk Assessment Guidance for Superfund (RAGS) (USEPA, 1989),
- Calculating Upper Confidence Limits for Exposure Point Concentrations at Hazardous Waste Sites (USEPA, 2002a),
- Supplemental Guidance for Developing Soil Screening Levels for Superfund Sites (USEPA, 2002b),
- Exposure Factors Handbook (EFH) (USEPA, 1997),
- RAGS Part E, Supplemental Guidance for Dermal Risk Assessment (USEPA, 2004a), and
- Tiered Approach to Corrective Action Objectives (TACO) (Illinois EPA, 2001).

Specifically, this risk assessment is intended to satisfy the requirements of 35 IAC 742.915 for the preparation of formal risk assessments. The formal HHRA is based specifically upon USEPA RAGS guidance and other pertinent documentation as indicated above. Information regarding sampling and analysis, extent of Tier 1 RO exceedances, and characteristics of the COCs are provided in the letter report "Project Category 4: Additional Specialized

Environmental & Engineering Services, Former Celotex Site – 2800 S. Sacramento Avenue, Chicago, Illinois' (URS, 2006).

The main components of the risk assessment are as follows:

- Exposure assessment;
- Toxicity assessment;
- Risk characterization; and
- Uncertainty Evaluation.

This formal risk assessment represents a site-specific risk assessment relating to current and potential future land use scenarios by using the following parameters:

- Conservative default exposure parameters;
- Site-specific exposure data relative to current and future land use; and
- Site-specific soil physical properties

4.1 Exposure Assessment

The following components must exist for an exposure pathway to be complete:

- A source and mechanism of chemical release;
- A retention or transport medium;
- A point of potential human contact with the impacted medium; and
- An exposure route at the contact point.

If one or more of these components are absent, the exposure pathway is incomplete. A potentially complete exposure pathway is one in which one or more of the four components is currently absent, but may become present under some future scenarios. In this formal risk assessment, the exposure assessment was conducted for both complete and potentially complete exposure pathways at the Site (Section 4 of Attachment A).

The Site is being evaluated based on guidance provided by CDOE indicating that the future land use of the Site will be recreational. Therefore, potential receptors include adult and child recreational users. Based on the current land use of industrial and anticipated future land use of recreational, and presence of sensitive populations, such as the elderly or small children, the land use selected to model risks at the Site is the residential land use scenario. The exposure routes evaluated for residential receptors are ingestion, inhalation of cPAHs in gravel fines, and dermal contact with cPAHs in gravel fines.

Although there are current industrial/commercial workers and it is possible for trespassers to access the Site, these exposure scenarios were not evaluated for this HHRA. The risks calculated

for the residential scenario are based on the most conservative exposure assumptions for all current and future receptors to the Site. Therefore, the residential scenario is considered protective of current industrial/commercial workers and potential trespassers as well as future recreational receptors.

For the purpose of this formal HHRA, residential receptor exposure was evaluated using a combination of default exposure parameter values to represent a conservative, yet reasonably site-specific exposure scenario. Some of the parameters, e.g., the 95% UCL of the mean, are meant to represent high-end exposure by using upper-bound estimates for exposure parameter values. The site-specific parameter values represent more reasonable exposure conditions that are applicable to the current and future land use scenarios.

4.1.1 Exposure Point Concentrations

The exposure point concentration (EPC) for each of the seven cPAHs was derived through statistical analysis of the data collected from the Site. In total, 59 sample results were used for calculating cPAH EPCs.

A statistical analysis was conducted using USEPA's ProUCL software (USEPA, 2004b), which follows guidance for calculating UCLs presented in *Calculating Upper Confidence Limits for Exposure Point Concentrations at Hazardous Waste Sites* (USEPA, 2002a). The EPC is generally based on the 95% UCL of the arithmetic mean. Different statistical methods were applied to calculate the 95% UCL based on sample size, percentage of detections, and distribution of the data set.

- In general, if the frequency of detection was less than 50%, non-parametric statistical methods were used to estimate the EPC. If the frequency of detection was greater than 50%, the raw data or log-normally transformed data were tested for normality and an appropriate method was used to estimate the EPC.
- For non-detect results, one-half of the reporting limit was used as the concentration in the statistical calculation (USEPA, 1992).
- When both original and field duplicate sample results were available, the average value of the original and duplicate data was used to represent the constituent concentration for a given location and depth.

The analytical data for the gravel fines samples collected for this investigation were provided in URS, 2006. A figure depicting the locations of the samples collected at the Site is also provided as Figure 1.

A summary of EPCs calculated for the HHRA is provided in Table 2. The output files for the EPC calculations for gravel fines are presented in Attachment B, Tables B-2 to B-8.

4.1.2 Estimating Chemical Intakes

The equations used to estimate exposure intakes of cPAHs were obtained by solving the Soil Screening Level (SSL) equations for the target risk term as set forth in TACO, Appendix C. Sources of the exposure factors include recommended default values from TACO. For calculation of dermal contact risk, Equations 3.11, 3.12, 3.21, 4.2, and 5.1 from RAGS Part E (USEPA, 2004a) were used.

To be conservative, chemical intakes were estimated using the 95% UCL of the mean as the EPC and TACO default values for exposure frequency, exposure duration, body weight and averaging time.

Exposure parameter values and equations used to estimate risk are presented in Tables 3-7 for the residential scenario. For comparative purposes, site-specific exposure parameter values for the recreational scenario are presented in Tables 9-11.

4.1.3 Toxicity Assessment

Chemical intake estimates are combined with descriptors of the chemical's potential toxicity, referred to as toxicity values. The result is an estimate of potential health risks associated with the exposure. Only carcinogenic toxicity is considered in this formal risk assessment for cPAHs, as noncarcinogenic toxicity values, e.g., reference doses (RfDs), have not been derived.

The toxicity value describing potential carcinogenicity of a chemical is called a cancer slope factor (SF) expressed in the units of $(mg/kg-day)^{-1}$. An SF represents an upper bound estimate of the probability that an individual may develop cancer following exposures to the particular chemical. Toxicity values were obtained from the following sources.

- 1. Integrated Risk Information System (IRIS) on-line database (USEPA, 2006)
- 2. Provisional toxicity values obtained from National Center for Environmental Assessment (NCEA), as published in USEPA Region 9 Preliminary Remediation Goals (PRGs) (USEPA, 2004c)

Toxicity values and chemical-specific values used in the risk calculations are provided in Table 3.

4.2 Risk Characterization

Cancer risks are expressed as the excess probability of cancer as a result of chemical exposure, and were estimated by multiplying the chemical intake by the SF, or Risk = Intake \times SF. Chemical-specific cancer risks were then summed for each pathway to yield a total excess risk for carcinogenic effects.

Acceptability of the overall cancer risk is typically gauged by comparing the risk estimate with the risk range of 1×10^{-4} to 1×10^{-6} (excess cancer risks of one in ten thousand to one in one

million). This risk range is the target risk level established by the National Oil and Hazardous Substances Contingency Plan for evaluating the need for and the extent of remediation (USEPA, 1990). Remediation is typically not warranted if the cumulative site risk is within this range.

The total cancer risk was estimated for potential future residential receptors exposure through the ingestion and inhalation of cPAHs in gravel fines at the Site (Tables 4 and 5). The calculated total cancer risks were 5×10^{-5} (Table 8). This risk estimate for exposures to gravel fines is above the TACO default risk level of 1×10^{-6} for no further action and within the acceptable risk range of 1×10^{-4} to 1×10^{-6} established by both IEPA and USEPA. However, it must be noted that risk was calculated for the more conservative residential land use scenario and risks associated with surficial exposure for the planned future recreational land use scenario would be much lower. Tables 9-11 present the risks to an adolescent (6-18 years of age) recreational receptor exposed to cPAHs in gravel fines. Risk for this receptor is 2×10^{-6} , which is just slightly above the lower end of the target risk range used by both the IEPA and USEPA.

4.3 Uncertainty Analysis

The formal risk assessment for current and future residential exposures is a conservative assessment of potential health risks posed by cPAHs in soil. The primary sources of uncertainty are discussed below.

It has been widely recognized by USEPA that repeated use of upper bound values for exposure parameters could lead to a substantial overestimate of the actual risk. Researches have reported that this approach could yield risk estimates for individuals that lie well above the intended 95th percentile (Finley et al., 1993). This conservative approach can readily lead to unnecessary overprotection and misplacing cleanup activities. This risk assessment used the default values and therefore, more conservative values for exposure frequency (EF) and exposure duration (ED). Based on the assumptions for exposure frequency and exposure duration for recreational receptors, site-specific exposure patterns are as much as one-half of IEPA's default values. Estimates of risk using site-specific information include the use of an ED of 12 years and an EF of 52 days per year results in a risk estimate of 2×10^6 , which is within both the IEPA and USEPA target risk range¹. Calculations using site-specific values are presented in Tables 8 to 10.

The assumptions made regarding the degree to which exposure occurs are conservative. The upper bound estimates of detected concentrations in surficial fill materials (i.e., 95% UCLs of the mean) were used as EPCs in the risk evaluation. The use of the 95% UCL of the mean is consistent with risk assessment guidance and represents an upper bound value; thus, this approach contributes to the conservatism in the overall evaluation.

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¹ The recreational adolescent exposure duration is assumed to range in age from 7 to 18. Therefore, total exposure duration is 12 years. The exposure frequency is a conservative assumption which assumes adolescents will be present 3 days per week during June, July, and August and 1 day per week during April, May, September, and October).

There are no published dermal SFs available for any chemicals in any USEPA database. However, based on literature evidence, cPAHs have been shown to induce systemic toxicity and tumors at distant organs. For this reason, the lack of a dermal toxicity value may not accurately predict risk for receptors exposed to cPAHs. Therefore, *RAGS Part E* (USEPA, 2004), only recommends a qualitative evaluation of the carcinogenic effects of PAHs. Although a quantitative evaluation was completed for this HHRA, the actual risks associated with this exposure pathway are unknown.

The oral and inhalation toxicity values applied in this risk assessment were derived by USEPA. The methodology by which toxicity values are derived is intentionally conservative. Use of the toxicity values has likely resulted in an overestimation of potential health risks.

4.4 Formal Risk Assessment Conclusions

A formal risk assessment for the cPAHs detected in gravel fines at the Site was conducted in accordance with USEPA risk assessment methodologies and IEPA's TACO regulations, which are based upon USEPA methodologies. Site-specific risk estimates were calculated for off-site residential receptors exposed to cPAHs in gravel fines. The conservative risk estimate associated with this receptor was calculated to be 5×10^{-5} . This estimate is above the default risk level of 1×10^{-6} used in TACO for no further action but within the risk range of 1×10^{-4} to 1×10^{-6} considered to be acceptable by USEPA and IEPA. Site-specific calculations of risks associated with recreational exposure to cPAHs in gravel fines is 2×10^{-6} which is slightly above the IEPA default risk level for no further action but within USEPA target risk range for no further action.

Therefore, based on the findings of the HHRA, the potential adverse health effects associated with exposures to concentrations of cPAHs in gravel fines are within acceptable levels and the Site is acceptable for recreational land use.

5.0 SUMMARY OF RISK EVALUATION

The risk evaluation for the Site was conducted in accordance with 35 IAC 742. The results of the risk evaluation are summarized below.

- No further action is needed for detections of cPAHs in gravel fines. The Tier 3 evaluation indicates the risks associated with exposures to these chemicals in gravel fines are within acceptable limits.
- No further action is needed for methylene chloride, cPAHs, dieldrin, and chromium in gravel fines. Based on the existence of the City of Chicago groundwater ordinance preventing the use of groundwater beneath the Site and the fate and transport characteristics for these chemicals, impacts to off-site residential receptors is nonexistent.

5.0 REFERENCES

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Table 1
Chemical Concentrations Exceeding TACO Tier I ROs
Former Celotex Site - 2800 S. Sacramento Ave.
Chicago, IL

Analyte	Chicago Background Levels	Specific	ntial Route Values for Soil	Route Spe	ion Worker ecilic Values Soil	Groun Ingestion	ponent of dwater Exposure Values		S2	S3	\$4	S 5	S6	S7	S8	S 9	<u>\$1</u> 0	S11
Volatile Organic Compounds		Ingestion	Inhalation	Ingestion	Inhalation	Class	Class !!											
Methylene chloride	NA	85	13	12,000	34	0.02	0.2	0.038 UJ	0.021 UJ	0.017 UJ	0.04 UJ	0 056 UJ	0.045 UJ	0.018 UJ	0 023 UJ	0 021 UJ	0.027 UJ	0 022 UJ
Semivolatile Organic Compounds																		
Benz(a)anthracene	11	0.9	NE	170	NE	2	8	1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1	* 1.8	6.2	1 3	Mat. 14 7 362	0.87	3.2.2.T	3.8	2011 A. 1600 C.	1,2	36.31.9a
Benzo(a)pyrene	13	0.09	NE	17	NE	8	82	4.9	1. 2.2 - FL	6.3	2.6	0.89 J	0.59	4 1.6 × 1.6 %	5 45 Te	\$ 1.2	- 0.88	12
Benzo(b)fluoranthene	1.5	0.9	NE	170	NE	5	25	5.3	2.6	, 7.1	3.2	121	0 72	2	4.7	1.7	1.1	1.7
Benzo(k)fluoranthene	1	9	NE	1,700	NE	49	250	3	1.2	6.4	1.6	061 J	0.43	0.96	32	0 74	0.7	0 69
Chrysene	11_	88	NE	17,000	NE	160	800	37	19	59	2.7	1 J	0 68	21	4	1.9	1.3	1.8
Dibenz(a,h)anthracene	02	0 09	NE	17	NE	2	7.6	0.96	0.49	1	0.55	0.19 J	0.12	0.32	0.62	0.24	0.13	0.24
Indeno(1,2,3-cd)pyrene	0.86	0.9	NE	170_	NE	14	69	3.5	1.3	3.7	91.4	0 38 J	0 32	0.83	2.8	0 67	0 52	0 65
Metals														_				
Chromium	16.2	230	270	4,100	690	21-36	NE	12	16	17	14	14	13	_16	13	13	14	12
Pesticides																		
Dieldrin	NA	0.04	1	78	3.1	0.004	0.2	0 0033 U	0.0032 U	0.0032 U	0 0033 U	0 0032 U	0 0033 U	0.0072	0.0033 U	0.005	0.0033 U	0 0033 U

Table 1
Chemical Concentrations Exceeding TACO Tier | ROs
Former Celotex Site - 2800 S. Sacramento Ave.
Chicago, IL.

Analyte	Chicago Background Levels	Specific	itial Route Values for Soil	Route Spe	ion Worker cific Values Soil	Grour Ingestion	nponent of ndwater Exposure Values]	S13	S14	S15	S16	S17	S18	S19	S20
Volatile Organic Compounds		Ingestion	Inhalation	Ingestion	Inhalation	Class I	Class II									
Methylene chloride	NA	85	13	12,000	34	0.02	0.2	0 084 UJ	0.056 UJ	0.079 UJ	0 02 UJ	0.05 UJ	0.052 UJ	0.039 UJ	0.026UJ	0.06 UJ
Semivolatile Organic Compounds										-						
Benz(a)anthracene	1.1	0.9	NE	170	NE	2	8	1 gl/4g (1966)	0.86	0.67	14. 12. 6 - 6 fg	0.67	~ 2.5	1,31	60, 9,9 0	24
Benzo(a)pyrene	1.3	0.09	NE	17	NE	8	82	0.86	~ 0.58 × · ·	0.48	0.72	0.42	1801466653	0.745	#6:07(b) (*6.3)	1.7.1.7 (C)
Benzo(b)fluoranthene	1.5	0.9	NE	170	NE	5	25	9 -1.3.0 has	0.85	0.72	- 1.1 A Sec.	0.57	> 2.1 Cesar	al dine	8.7	A. 21年の
Benzo(k)fluoranthene	1	9	NE	1,700	NE	49	250	0.53	0.3	0.26	0.4	0.32	0.89	0.43 J	4.4	1.3
Chrysene	1.1	88	NE	17,000	NE	160	800	1.4	0.88	0.72	1.2	0.67	2.4	1.3 J	8.6	2.3
Dibenz(a,h)anthracene	0.2	0.09	NE	17	NE	2	7.6	0.12	0 076	0.11	0.099	0.1	0.33	0173	i 0.91	A. 0.42
Indeno(1,2,3-cd)pyrene	0.86	0.9	NE	170	NE	14	69	0 49	0.33	0.28	0.4	0.24	0.88	0.46 J	18 3 H 1 1 1 1 1	0.92
Metals																
Chromium	16.2	230	270	4,100	690	21-36	NE	16	17	13	15	12	19	17	16	19
Pesticides																
Dieldrin	NA	0.04	1	7.8	3.1	0.004	0.2	0.0033 U	0 0032 U	0.0043	0 0033 U	0.0033 U	0.0053	0.0033 U	0.0033 U	0.0046

Table 1
Chemical Concentrations Exceeding TACO Tier I ROs
Former Celotex Site - 2800 S. Sacramento Ave.
Chicago, IL

Analyte	Chicago Background Levels	Specific \	tial Route Values for oil	Route Spe	ion Worker cific Values Soil	Groun Ingestion	nponent of idwater Exposure Values		S22	S23	S24	\$25	S26	S-27	S-28	S-29	S-30	\$-3 1
Volatile Organic Compounds		Ingestion	Inhalation	Ingestion	Inhalation	Class I	Class II							I				
Methylene chloride	NA NA	85	13	12,000	34	0.02	0.2	0 029 W	0 042 UJ	0 022 UJ	0.049 UJ	0.046 UJ	0 033 ÜÜ	0.04 W	0 055 UJ	0 029 UJ	0 073 UJ	0.038 UJ
Semivolatile Organic Compounds																		
Benz(a)anthracene	1.1	0.9	NE	170	NE	2	8	0.84	17 1.1 Tal 18 18 18 18 18 18 18 18 18 18 18 18 18	0 25	3613 C. C.	2.2	0.52	29	H. 13.14	0.79 J	3.1.1.25 mile	25.45.000
Benzo(a)pyrene	1.3	0 09	NE	17	NE	8	82	0.66	0.85	0.22	13	1,9	0 44	23.	Post Market	0.84 J	· 电影子的	22 THE
Benzo(b)fluoranthene	1.5	0.9	ΝE	170	NE	5	25	0.85	12	0.29	1.3	2.3	0.56	3.1	1.2	0.89 J	1.3	2.7
Benzo(k)fluoranthene	1	9	NE	1,700	NE	49	250	0.5	0.75	0.14	1	1.4	0.43	12	0.69	034 J	0.83	1.5
Chrysene	1.1	88	NE	17,000	NE	160	800	0.86	1.1	0.25	1.3	2.2	0.54	29	1.3	0.78	1.2	22
Dibenz(a,h)anthracene	0.2	0.09	NE	17	NE	2	7.6	0.14	0.19	0.073	0.26	0.38	0.12	0.32	1 0.19 Land	0.13	0.14: Links	0.42
Indeno(1,2,3-cd)pyrene	0.86	0.9	ΝE	170	NE	14	69	0.31	0 47	0.14	0 63	0.95	0.24	1.1	0 49	0.33 J	0.57	25-13 Sept. 1989
Metals	T																	
Chromium	16.2	230	270	4,100	690	21-36	NE	13	14	11	17	17	18	16	16	17	19	25 55
Pesticides																		
Dieldrin	NA NA	0 04	1	7.8	3.1	0 004	0.2	0.0033 U	0.006	0 0033 U	0.0034	0 0036	0 0032 U	0.0037	0.0033 U	0 0033 U	0 0033 U	0.0063

Table 1
Chemical Concentrations Exceeding TACO Tier I ROs
Former Celotex Site - 2800 S. Sacramento Ave.
Chicago, II.

Analyte Volatile Organic Compounds	Chicago Background Levels	Specific S	Soil	Route Spe for	Soil	Ingestion Route	dwater Exposure Values)	S-33	S-34	S-35	S-36	S-37	S-38	S-39	S-40	S-41_	S-42
Methylene chloride	NA NA	Ingestion 85	Innaiation	Ingestion 12,000	nnalation	0.02	0.2	0.038 UJ	0.068 UJ	0 06 UJ	0 059 UJ	0.062 ÚJ	0 C59 UJ	0 052 UJ	0.041 UJ	0.034 UJ	0.061 UJ	0.037 LJ
Semivolatile Organic Compounds	192	- 03	1 13 -	12,000	- 34	0.02	0.2	0.036 00	0 000 00	000 00	0.009 00	0.002 03	0 000 00	0 032 03	0.041 03	0.034 03	0.001 03	0.007
Benz(a)anthracene	1.1	09	I NE	170	NE	1 2	8	2.2	0.03	0.89	0.35 J	058 3 467-56	F-11 F-3	0.77	1. The	E-S1483565	₹19. (%)	0.6 J
Benzo(a)pyrene	1.3	0.09	NE	17	NE.	8	82	*X15	(0.00 ± 70	0.77	0.31 J	-28	0.39	*. 0.7	12.6			0.45
Benzo(b)fluoranthene	15	0.9	NE	170	NE	5	25	7.200 Des	1.13	0.92	0 39 J	13	0.66	0.69	4.17		1.8	0.58 J
Benzo(k)fluoranthene	1	9	NE	1,700	NE	49	250	0.92	0.75	0.67	0.24 J	12	0.48	0 67	0.94	0.82	0.85	0.33 J
Chrysene	1.1	88	NE	17,000	NE NE	160	800	2	1	0.65	04 J	2.9	1.2	0.84	2	13	2	0.61 J
Dibenz(a.h)anthracene	0.2	0.09	NE	17	NE	2	76	0.32	0.22	0.19	0 084 J	0.38	0.15	0.084	0.22	0.24	0.28	0.093
Indeno(1,2,3-cd)pyrene	0.86	0.9	NE	_170	NE	14	69	0.78	0.5	0 44	02J	1.4	0.28	0.39	0.61	0 48	0.73	0.25 J
Metals	l																	
Chromium	16.2	230	270	4,100	690	21-36	NE	17	16	16	16	11	17	14	18	22:00	2 21 7	15
Pesticides																		
Dieldrin	NA	0.04	1	7.8	3.1	0.004	0.2	0 0032	0 0033 U	0 0033 U	0 0033 U	0.0055	0.0057	0.0073	0.0046	0.0047	0.005	0 0033 U

Table 1
Chemical Concentrations Exceeding TACO Tier I ROs
Former Celotex Site - 2800 S. Sacramento Ave.
Chicago, IL

Analyte	Chicago Background Levels	Specific S	Soil	Route Spe for	ion Worker cific Values Soil	Ingestion Route	dwater Exposure Values]	S-44	S-45	S-46	S-47	S-48	5-49	S-50	<u>S-51</u>	S-52	S-53	S-54
Volatile Organic Compounds	T	Ingestion	Inhalation	Ingestion	Inhalation	Class !	Class II												
Methylene chloride	NA .	85	13_	12,000	34	0.02	0.2	0.05 UJ	0,036	0.028	0.018 J	0.034 1	0.057 J	0.034 4	0.048 J	0.028 J	0.05 J	0.025 J	0.059
Semivolatile Organic Compounds																			
Benz(a)anthracene	1.1	0.9	NE	170	NE	2	8	1. 1.3 Table 3. 15	4 MG - 1	。 (14)	35 1.35 FE	0.87	0.84	0.78	(名,1.1元年)40;	1.1 4 5	25.11.82	0.65	0.64
Benzo(a)pyrene	1.3	0.09	NE	17	NE	8	82	0.55	0.9.4	0.77	2.0,86	0.9	0.81	0.82	0.79	- 0.8	0.53	0,39	0.41
Benzo(b)fluoranthene	1.5	0.9	NE	170	NE	5	25	0.79	1.2	1.4	1144	1.2	1981.11 1985.	0.76	[10 4 [17]]	1.2	0.75	0.49	0.56
Benzo(k)fluoranthene	1	9	NE	1,700	NE	49	250	0.58	0.79	0.68	0.85	0.54	0.46	0.81	0.7	0.92	0.53	0.43	0.37_
Chrysene	1.1	88	NE	17,000	NE	160	800	1.4	1.1	12	15	0 92	0.8	0.86	1	1.1	12	0.68	0.65
Dibenz(a,h)anthracene	0.2	0.09	NE	17	ΝĒ	2	7.6	0.19	0.19	0.10	0 24	0.22	0.18	0.25	0 12	0.17	0.18	0.11	0.11
Indeno(1,2,3-cd)pyrene	0.86	0.9	NE	170	NE	14	69	0.33	0 52	0.34	0 47	0.52	0 45	0.5	0.39	04	0 33	0 26	0 27
Metals																			
Chromium	16.2	230	270	4,100	690	21-36	NE	14	16	16	15	16	15	16	14	15	17	15	13
Pesticides																			
Dieldrin	NA	0.04	1	7.8	3.1	0.004	02	0.0083	0.0053	_ 0.0033 U	0.0033 ರ	0.0033	0 0033 U	0.0036	0.0073	0 0033 U	0 0033 U	0.0032 U	0.0033

Table 1 Chemical Concentrations Exceeding TACO Tier I ROs Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

<u> </u>			-			Soil Con	nponent of				_	T
	Chicago	Residen	tial Route	Construct	ion Worker	Grout	ndwater				1	
	Background	Specific 1	Values for	Route Spe	cific Values	Ingestion	Exposure				ĺ	
Analyte	Levels	S	ioil	for	Soil	Route	Values	S-55	S-56	S-57	S-58	S-59
Volatile Organic Compounds		Ingestion	Inhalation	Ingestion	Inhalation	Class I	Class II					
Methylene chloride	NA	85	13	12,000	34	0 02	0.2	0.068 J	0.074 × J	0.031 3	் 0.059 ். J .	0.094 J
Semivolatile Organic Compounds												
Benz(a)anthracene	1.1	0.9	NE	170	NE	2	8	1245	35147 AV	1.61	0.77	Lionali
Benzo(a)pyrene	1.3	0.09	NE	17	NE	8	82	0.92-0	7 0.76	0.713	0.6	0.99
Benzo(b)fluoranthene	1.5	0.9	NE	170	NE	5	25	2014年11月1日	1.0	0.83	0.78	141.4 mm
Benzo(k)fluoranthene	1	9	NE	1,700	NE	49	250	0.48	0.68	0.59	06	0.76
Chrysene	1.1	88	NE	17,000	NE	160	800	1	1.4	17J	0 79	1.2
Dibenz(a,h)anthracene	_0.2	0.09	Z	17	NE .	2	7.6	0.18	一0.190	0.26	0.2	0.22
Indeno(1,2,3-cd)pyrene	0.86	0.9	ΝE	170	NE	14	69	0.5	0.37	0.5	0.36	0.54
Metals												
Chromium	16.2	230	270	4,100	690	21-36	NE	12	11	16	13	14
Pesticides												
Dieldrin	NA	0.04	1	7.8	3.1	0.004	0.2	0.0036	0.0032 U	6.01	0 0033 U	0.0033 U

Table 1 Chemical Concentrations Exceeding TACO Tier I ROs Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Notes:

Concentrations are in milligrams per kilogram (mg/kg)	
Bold and shaded values exceed Migration to Class I Groundwater ROs.	
Black Bold and shaded values exceed Residental and/or Construction Worker ROs.	
alicized values exceed Residential and or Construction Worker ROs but are below Chicago Background Levels.	
/alues shown within a bold cell. = also exceed Migration to Class I Groundwater ROs.	
Ranges for certain Migration to Groundwater ROs are pH-based according to TACO Section 742 Appendix B, Tab.	ile C
NIP NICE CONTROL OF	

- NE Not Established
- NA Not Analyzed
 - J Indicates an estimated concentration because of results below the sample reporting limit, or results where QC criteria were not met
- UJ Indicates that the analyte was not detected at or above the sample reporting limit. However, because of QC issues, the reporting limit is approximate and may or may not represent the actual limit of reporting necessary to accurately and precisely measure the analyte in the sample.
- R Result Rejected. Presence or absence of compound cannot be determined

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Printed: 2/7/2007

Table 2 Exposure Point Concentration - Gravel Fines Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Analyte	Maximum Concentration (mg/kg)	Average Concentration (mg/kg)	95% UCL (mg/kg)	Distribution	Statistic used for 95% UCL	Rationale	Selected EPC (mg/kg)
Benz(a)anthracene	9.90E+00	1.62E+00		non-parametric	Use 95% Chebyshev (Mean, Sd) UCL	UCL <max< td=""><td>2.46</td></max<>	2.46
Benzo(a)pyrene	7.00E+00	1.30E+00	2.06E+00	non-parametric	Use 95% Chebyshev (Mean, Sd) UCL	UCL <max< td=""><td>2.06</td></max<>	2.06
Benzo(b)fluoranthene	8.70E+00	1.62E+00	2.48E+00	non-parametric	Use 95% Chebyshev (Mean, Sd) UCL	UCL <max< td=""><td>2.48</td></max<>	2.48
Benzo(k)fluoranthene	6.40E+00	9.45E-01	1.52E+00	non-parametric	Use 95% Chebyshev (Mean, Sd) UCL	UCL <max< td=""><td>1.52</td></max<>	1.52
Chrysene	8.60E+00	1.60E+00	1.84E+00	non-parametric	H-UCL	UCL <max< td=""><td>1.84</td></max<>	1.84
Dibenz(a,h)anthracene	1.00E+00	2.54E-01	2.91E-01	parametric	H-UCL	UCL <max< td=""><td>0.29</td></max<>	0.29
Indeno(1,2,3-c,d)pyrene	3.70E+00	7.21E-01	1.15E+00	non-parametric	Use 95% Chebyshev (Mean, Sd) UCL	UCL <max< td=""><td>1.15</td></max<>	1.15

Notes:

All units are milligrams per kilogram (mg/kg). EPC = Exposure Point Concentration

UCL = Upper Confidence Limit

MVUE = Minimum Variance Unbiased Estimator

NA = Not Applicable

Table 3 Chemical-Specific and Toxicity Values Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Chemical		Solubility in Water S (mg/L)		Dimensionle Henry's Lav Constant H' (unitless)	~	Organic Carbon:Wat Partition Coefficient Koc (cm ³ /g)	ər	Diffusion Coefficient in Air D _{alr} (cm²/s)	Diffusion Coefficier in Water D _{waler} (cm ² /s)	- 1	Gastrointestinal Absorption Efficiency ABSgi (unitless)	Dermal Absorption Fraction ABSd unitless	on 1	Chronic Oral Reference Dose RfD ₆ [2] (mg/kg-dy)	Chronic Inhalation Reference Concentration RfC [2] (mg/m ³)	Oral Slope Factor SF, (mg/kg-dy)	Inhalation Slope Factor SF _i (mg/kg-dy) ⁻¹	Inhalation Unit Risk UR, (μg/m ³)·¹
Benzo(a)pyrene Benzo(b)fluoranthene Benzo(k)fluoranthene Chrysene Dibenz(a,h)anthracene	Car Car Car Car Car	9.40E-03 [1 1 62E-03 [1 1.50E-03 [1 8.00E-04 [1 1 60E-03 [1	j 	1.37E-04 4.63E-05 4.55E-03 3.40E-05 3.88E-03 6.03E-07 6.56E-05	[1] [1] [1] [1] [1] [1]	3.98E+05 1.02E+06 1.23E+06 1.23E+06 3.98E+05 3.80E+06 3.47E+06	[1] [1] [1] [1] [1] [1]	5.10E-02 [1] 4.30E-02 [1] 2.26E-02 [1] 2.26E-02 [1] 2.48E-02 [1] 2.02E-02 [1] 1.90E-02 [1]	9 00E-06 [1 5.56E-06 [1 5.56E-06 [1 6.21E-05 [1 5.18E-06 [1	i	1 00E+00 [2] 1.00E+00 [2] 1.00E+00 [2] 1 00E+00 [2] 1 00E+00 [2] 1 00E+00 [2] 1.00E+00 [2]	1 30E-01 1 30E-01 1 30E-01 1 30E-01 1 30E-01 1 30E-01	[3] [3] [3] [3] [3] [3]	NA NA NA NA NA NA	NA NA NA NA NA NA	7.30E-01 NCEA 7.30E-00 IRIS 7.30E-01 NCEA 7.30E-02 NCEA 7.30E-03 NCEA 7.30E-00 NCEA 7.30E-01 NCEA	3.1 NCEA NA IRIS NA IRIS NA IRIS NA IRIS	NA IRIS 8.86E-04 NCEA NA IRIS NA IRIS NA IRIS NA IRIS

Notes:

[1] Appendix C, Table E of TACO (Illinois EPA, 2001).

[2] USEPA, 2004. RAGS, Part E, Exhibit 4-1. The % Absorbed ABS₆ is greater than 50% RAGSE recommends no adjustment for chemicals in this category.

[3] USEPA, 2004. RAGS, Part E, Exhibit 3-4. Recommended dermal absorption fraction from soil benzo(a) pyrene and other PAHs

IRIS = U.S. EPA's Integrated Risk Information System (http://www.epa.gov/iris)

Car = Carcinogen

NCEA = EPA-NCEA Regional Support provisional value

NA = Not available.

Table 4 Risk Calculation for the Ingestion Route for Gravel Fines Residential Land Use Former Celotex Site - 2800 S, Sacramento Ave. Chicago, IL

Residential Land Use - Gravel Fines

Equation:	$Risk_{soiling} = Csoil \times SF_o \times CF_2 \times EF \times IFsoil-adj$		
	$AT_c \times CF_1$		
where:	C _{soil} = Concentration in Soil (mg/kg)	Calculated, 95% UCL	(See Table 2)
	$SF_o = Oral Slope Factor (mg/kg.day)$	Chemical-specific	(See Table 3)
	CF ₂ = Unit conversion Factor (kg/mg)	1E-06	
	EF = Exposure frequency (dy/yr)	350	TACO Default for residential land use
	IF _{soil-ad} = Age-adjusted Soil Ingestion Factor for Carcinogens (mg-yr/kg-d)	114	TACO Default for residential land use
	BW≃ Body Weight (kg)	70	TACO Default
	$AT_c = Averaging Time (yr)$	70	TACO Default
	CF_1 = Unit conversion factor (dy/yr)	365	

Chemical	Csoil	SF _o	CF ₂	EF	IFsoil-adj	AT_{c}	CF _t	Risk
	(mg/kg)	(mg/kg.day) ⁻¹	(kg/mg)	(dy/yr)	mg-yr/kg-d	(yr)	(dy/yr)	
Benz(a)anthracene	2.46	7.30E-01	1.0E-06	350	114	70	365	2.80E-06
Benzo(a)pyrene	2.06	7.30E+00	1.0E-06	350	114	70	365	2.35E-05
Benzo(b)fluoranthene	2.48	7.30E-01	1.0E-06	350	114	70	365	2.83E-06
Benzo(k)fluoranthene	1.52	7.30E-02	1.0E-06	350	114	70	365	1.73E-07
Chrysene	1.84	7.30E-03	1.0E-06	350	114	70	365	2.10E-08
Dibenz(a,h)anthracene	0.29	7.30E+00	1.0E-06	350	114	70	365	3.32E-06
Indeno(1,2,3-c,d)pyrene	1.15	7.30E-01	1.0E-06	350	114	70	365	1.31E-06
<u></u>							Total Pathway Risk	3E-05

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Table 5 Risk Calculations for Inhalation Route for Gravel Fines Residential Receptors Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Residential Land Use - Gravel Fines

Equation:	$Risk_{eoilinh} = Csoil \times URF \times CF_2 \times EF \times ED \times [$	(1/VF) + (1/PEF)]	_
	ÄT₀ × CF₁		
where:	C _{soil} = Concentration in Soil (mg/kg)	Calculated, 95% UCL	(See Table 2)
	URF ≈ Inhalation Unit Risk Factor((μg/m³) ⁻¹	See Table 3	
	$CF_2 = Unit conversion Factor (\mu g/mg)$	1000	
	EF = Exposure frequency (dy/yr)	350	TACO default
	ED = Exposure duration (yr)	30	TACO default
	VF ≈ Volatilization Factor (m³/kg)	calculated	
	PEF ≈ Particulate Emission Factor (m³/kg)	1.32E+09	TACO default for residential land use
	BW = Body Weight (kg)	70	TACO default
	$AT_c = Averaging Time (yr)$	70	TACO default
	CF_1 = Unit conversion factor (dy/yr)	365	

Chemical	Csoil	URF	CF ₂	EF	ED	VF	PEF	AT _c	CF,	Risk
	(mg/kg)	(μg/m³) ⁻¹	(μg/mg)	(dy/yr)	(yr)	(m³/kg)	(m³/kg)	(yr)	(dy/yr)	
Benz(a)anthracene	2.46	NA	1000	350	30	9.80E+06	1.32E+09	70	365	NA
Benzo(a)pyrene	2.06	8.86E-04	1000	350	30	1.57E+07	1.32E+09	70	365	4.82E-08
Benzo(b)fluoranthene	2.48	NA	1000	350	30	2.02E+07	1.32E+09	70	365	. NA
Benzo(k)fluoranthene	1.52	NA	1000	350	30	1.03E+06	1.32E+09	70	365	NA NA
Chrysene	1.84	NA	1000	350	30	4.01E+07	1.32E+09	70	365	NA NA
Dibenz(a,h)anthracene	0.29	NA	1000	350	30	4.01E+07	1.32E+09	70	365	NA
Indeno(1,2,3-c,d)pyrene	1.15	NA	1000	350	30	3.66E+07	1.32E+09	70	365	NA
	 							Tota	l Pathway Risk	5E-08

Table 6 Volatilization Factor for Inhalation Pathway Residential and Recreational Receptors Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Equation:

$$VF = \frac{Q}{C_{v_F}} \times \frac{(3.14 \times D_A \times T)^{1/2}}{(2 \times \rho_b \times D_A)} \times 10^{-4} \frac{m^2}{cm^2}$$

where:

VF = Volatilization Factor (m³/kg)

 Q/C_{VF} = Inverse of mean concentration at the center of 1 acre square source $(g/m^2-s)/(kg/m^3)$

 $\pi = pi (3.14)$

 $D_A = Apparent Diffusivity (cm²/s)$

T = Exposure interval (s), 30 yrs for residential receptors

 ρ_b = Dry soil bulk density (g/cm³), site-specific

Chemical	Q/C _{VF} (g/m ² -s)/(kg/m ³)	π	D _A (cm²/s)	T (s)	2	₽ _b (g/cm³)	10 ⁻⁴ (m ² /cm ²)	VF (m³/kg)
Benz(a)anthracene	85.81	3,14	1.43E-10	9.50E+08	2	2	1,00E-04	9.80E+06
Benzo(a)pyrene	85.81	3.14	5.55E-11	9.50E+08	2	2	1.00E-04	1,57E+07
Benzo(b)fluoranthene	85,81	3.14	3.35E-11	9.50E+08	2	2	1.00E-04	2.02E+07
Carbazole	85.81	3.14	1.29E-08	9.50E+08	2	2	1.00E-04	1.03E+06
Dibenz(a,h)anthracene	85.81	3.14	8.55E-12	9.50E+08	2	2	1.00E-04	4.01E+07
Indeno(1,2,3-c,d)pyrene	85,81	3.14	1.03E-11	9.50E+08	2	2	1.00E-04	3.66E+07

Table 6
Volatilization Factor for Inhalation Pathway
Residential and Recreational Receptors
Former Celotex Site - 2800 S. Sacramento Ave.
Chicago. IL

Equation:

$$D_{A} = \frac{\left(\partial_{a}^{3,33} \bullet D_{i} \bullet H'\right) + \left(\partial_{w}^{3,33} \bullet D_{w}\right)}{\eta^{2}} \bullet \frac{1}{\left(\left(\rho_{b} \bullet K_{d}\right) + \partial_{w} + \left(\partial_{a} \bullet H'\right)\right)}$$

where:

 $D_{\Delta} = Apparent Diffusivity (cm²/s)$ calculated D_i = Diffusivity in Air (cm^{2/}s) chemical-specific Dw = Diffusivity in Water (cm²/s) chemical-specific θ_{a} Air-filled soil porosity (cm³/cm³) 0.05 θ_w= Water-filled soil porosity (cm³/cm³) 0.2 H' = Unitless Henry's Law constant chemical-specific $\eta = \text{Total soil porosity (cm}^3/\text{cm}^3)$ 0.25 ρ_b = Dry soil bulk density (g/cm³) 2 $K_d = K_{oc} \times f_{oc}$ calculated $K_{\infty} =$ Organic carbon partition coefficient (cm³/g) chemical-specific f_{oc} = Organic carbon content of soil (g/g) 0.006

Chemical	⊖ൂ (cm³/cm³)	⊖ _w (cm³/cm³)	D _i (cm²/s)	D _w (cm²/s)	H' (unitless)	η	ρ _δ (g/cm ³)	K _{ac} t _a (cm²(g) (grg)	K _d (cm³/g)	D _A (cm²/s)
Benz(a) anthracene	0.05	0.2	5.10 E- 02	9.00E-06	1.37E-04	0.25	2	3.98E+05 0.006	2.39E+03	1.43E-10
Benzo(a)pyrene	0.05	0.2	4.30E-02	9.00E-06	4.63E-05	0.25	2	1.02E+06 0.006	6.12E+03	5.55E-11
Benzo(b) fluoranthene	0.05	0.2	2.26E-02	5.56E-06	4.55E-03	0.25	2	1,23E+06 0,006	7.38E+03	3.35E-11
Benzo(k)fluoranthene	0.05	0.2	3.90E-02	7.03E-06	6.26E-07	0.25	2	3.39E+03 0.006	2.03E+01	1.29E-08
Chrysene	0.05	0.2	2.02E-02	5.18E-06	6.03E-07	0.25	2	3.80E+06 0.00B	2.28E+04	8.55E-12
Dibenz(a,h)anthracene	0.05	0.2	2.02E-02	5.18E-06	6.03E-07	0.25	2	3:80E+06 0:006	2.28E+04	8.55E-12
Indeno(1,2,3-c,d)pyrene	0.05	0.2	1,90E-02	5.66E-06	6.56E-05	0.25	2	3.47E+06 0.006	2.08E+04	1.03E-11

Table 7 Risk Calculation for Dermal Contact with Gravel Fines Residential Receptors Former Calotex Site - 2800 S. Sacramento Ave. Chicago, L

Residential Land Use - Gravel Fines

Equation: Risk_{soil-derm} = $\frac{(C_{sod} \times SF_{abs} \times CF_2 \times EF \times EV \times SCR_{adj} \times ABSd)}{AT_c \times CF_1}$

where:	C _{not} = Concentration in Soil (mg/kg)	Calculated, 95% UCL	See Table 2
	SF _{abs} = Slope Factor, absorbed (SF _o - ABS _g)	See Table 3	ABS _q , values are as presented in Exhibit 4-1 in RAGS Part E.
	CF₂ = Unit conversion Factor (kg/mg)	1E-06	
	EF = Exposure frequency (day/yr)	350	USEPA, 1989 Exhibit 6-14 of RAGS, Volume I, Human Health Evaluation Manual (Part A). Default exposure frequency
	SCR _{aq} = Age-adjusted Soil Contact Rate	360.28	Calculated using:
			$SCRadJ = \left(\frac{SAa \times EDa \times AFa}{BWa}\right) + \left(\frac{SAc \times EDc + AFc}{BWc}\right) = \left(\frac{570 cm^2 \times 24 y \times 0.07 mg/cm^2 - event}{70 kg}\right) \\ = + \left(\frac{2800 cm^2 \times 6 y \times 0.2 mg/cm^2 - event}{15 kg}\right) = 360.28 m^3 - y/kg - event$
	where.		
	SAc = Skin Surface Area Exposed, child (cm²)	2,800	USEPA, 2004 RAGS, Volume 1, Human Health Evaluation Manual (Part E) Exhibit 3-5. Default skin surface area
	SAa = Skin Surface Area Exposed, adult (cm²)	5,700	USEPA, 2004. RAGS, Volume 1. Human Health Evaluation Manuel (Part E) Exhibit 3-5. Default skin surface area.
	AFc = Soil-to-skin adherence factor, child (mg/cm²-event)	02	USEPA, 2004 RAGS, Volume 1, Human Health Evaluation Manual (Part E) Exhibit 3-5. Default soil-to-skin adherence factor.
	AFa = Soil-to-skin adherence factor, adult (mg/cm²-event)	0.1	USEPA, 2004. RAGS. Volume 1, Human Health Evaluation Manual (Part E) Exhibit 3-5 Default soil-to-skin adherence factor.
	EDc = Exposure Duration, Child (years)	6	USEPA, 1989 Exhibit 6-14 of RAGS, Volume I. Human Health Evaluation Manual (Part A).
	EDa ≂ Exposure Duration, adult (years)	24	USEPA, 1989 Exhibit 6-14 of RAGS, Volume I, Human Health Evaluation Manual (Part A)
	BWc = Body Weight, child (kg)	15	USEPA, 1989. Exhibit 6-14 of RAGS, Volume I. Human Health Evaluation Manual (Part A).
	BWa = Body Weight, adult (kg)	70	70 kg body weight and 70 year lifebme are used to be consistent with the development of cancer slope factors.
	EV ≈ Event Frequency (events/day)	1	USEPA, 2004 RAGS, Part E, Exhibit 3-5. Default event frequency.
	ABSd = Dermal Soil Absorption Factor (unitless)	chemical-specific	USEPA, 2004. RAGS. Part E, Exhibit 3-4. Recommended dermal absorption fraction from soil.
	$AT_c = Averaging Time (yr)$	70	TACO Default Value
	CF: = Unit conversion factor (days/year)	365	

Chemical	Caol	SF, ABS,	SF ₃₀₈	CF ₂	EF	EV	SCR	ABSd	ΑT _c	CF,	Risk _{eo⊪dem} ,
	(mg/kg)	(mg/kg:day) (unitless)	(mg/kg-day) 1	(kg/mg)	(day/yr)	(events/day)	(mg-yr/kg-event)	(unitiess)	(yr)	(day/yr)	
Benz(a)anthracene	2.46	7.30E-01 1.09E+00	7.30E-01	1.0E-06	350	1	360	1 30E-01	70	365	1.15E-06
Benzo(a)pyrene	2.06	7.00E+00 1.00E+00	7.30E+00	1.0E-06	350	1	360	1.30E-01	70	365	9.64E-06
Benzo(b)fluoranthene	2.48	7.30E-01 1.00E+00	7.30E-01	1,0E-06	350	1	360	1 30E-01	70	365	1.16E-06
Benzo (k)fluoranthene	1.52	7,30E-02 1.00E+00	7 30€-02	1.0E-06	350	1	360	1 30E-01	70	365	7.13E-08
Chrysene	1.84	7.30E-03 1.00E+00	7.30E-03	1.0E+06	350	1	360	1.30E-01	70	365	8.64E-09
Dibenz(a,h)anthracens	0.29	7.30E+00 1.00E+00	7 30E+00	1 0E-06	350	1	360	1.30E-01	70	365	1.36E-06
indeno(1,2,3-c.d)pyrene	1.15	7.30E-01 1.00E+00	7.30E-01	1.0E-06	350	1	360	_1.30E-01		365	5.37E-07
	1									Total Pathway Risk	1E-05

Table 8 Risk Characterization Summary Residential Land Use Former Celotex Site - 2800 S. Sacramento Ave. Chicago, JL

Risk Summary - Gravel Fines

Exposure Receptor	Risk Estimate for	Risk Estimate for Dermal	Risk Estimate for	Total Risk
	Ingestion Pathway	Pathway ¹	Inhalation Pathway	Estimate
Residential	3E-05	1E-05	5E-08	5E-05

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Table 9 Risk Calculation for the Ingestion Route for Gravel Fines Recreational Land Use Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Recreational Land Use - Gravel Fines

Equation:	$Risk_{soiling} = Csoil \times SF_o \times CF_2 \times EF \times IR \times ED$		
	AT _c × CF ₁		
where:	$C_{soil} = Concentration in Soil (mg/kg)$	Calculated, 95% UCL	(See Table 2)
	SF _o = Oral Slope Factor (mg/kg.day) ⁻¹	Chemical-specific	(See Table 3)
	CF ₂ = Unit conversion Factor (kg/mg)	1E-06	
	EF = Exposure frequency (dy/yr)	52	
	ED = Exposure duration (yr)	12	
	IF _{soil-adj} = Age-adjusted Soil Ingestion Factor for Carcinogens (mg-yr/kg-d)	100	
	BW= Body Weight (kg)	43	
	AT _c = Averaging Time (yr)	70	
	CF_1 = Unit conversion factor (dy/yr)	365	

Chemical	Csoil	SF ₀	CF ₂	EF	IR	ED ()	BW (fee)	AT _c	CF ₁	Risk
	(mg/kġ)	(mg/kg.day) ⁻¹	(kg/mg)	(dy/yr) 	(mg-yr/kg-d) 	(yr)	(kg)	(yr)	(dy/yr)	
Benz(a)anthracene	2.46	7.30E-01	1.0E-06	 52	100	12	43	70	365	1.02E-07
Benzo(a)pyrene	2.06	7.30E+00	1.0E-06	52	100	12	43	70	365	8.53E-07
Benzo(b)fluoranthene	2.48	7.30E-01	1.0E-06	52	100	12	43	70	365	1.03E-07
Benzo(k)fluoranthene	1.52	7.30E-02	1.0E-06	52	100	12	43	70	365	2.26E-08
Chrysene	1.84	7.30E-03	1.0E-06	52	100	12	43	70	365	2.74E-09
Dibenz(a,h)anthracene	0.29	7.30E+00	1.0E-06	52	100	12	43	70	365	1.21E-07
Indeno(1,2,3-c,d)pyrene	1.15	7.30E-01	1.0E-06	52	100	12	43	70	365	4.75E-08
									Total Pathway Risk	1E-06

Table 10 Risk Calculations for Inhalation Route for Gravel Fines Recreational Receptors Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Recreational Land Use - Gravel Fines

Equation: where:	Risk _{soilinh} =	Csoil × URF × CF ₂ × EF × ED × [(1/VF) + (1/PEF)] $AT_c \times CF_1$		
	C _{soil} = Concentration URE = Inhalation Uni	in Soil (mg/kg) t Risk Factor((μg/m³) ⁻¹	Calculated, 95% UCL See Table 3	(See Table 2)
	CF ₂ = Unit conversion	on Factor (μg/mg)	1000	
	EF = Exposure freq ED = Exposure dura	ation (yr)	52 12	
	VF = Volatilization F PEF = Particulate En	Factor (m³/kg) nission Factor (m³/kg)	calculated 1.32E+09	
	BW = Body Weight AT _c = Averaging Tim		43 70	
	CF ₁ = Unit conversion	* *	365	

Chemical	Csoil	URF	CF ₂	ÉF	ED	VF	PEF	ΑΤ _c	CF₁	Risk
	(mg/kg)	$(\mu g/m^3)^{-1}$	(µg/mg)	(dy/yr)	(yr)	(m ³ /k g)	(m³/kg)	(yr)	(dy/yr)	
Benz(a)anthracene	2,46	NA	1000	52	12	9.80E+06	1.32E+09	43	365	NA
Benzo(a)pyrene	2.06	8.86E-04	1000	52	12	1.57E+07	1.32E+09	43	365	4.66E-09
Benzo(b)fluoranthene	2.48	NA	1000	52	12	2.02E+07	1.32E+09	43	365	NA NA
Benzo(k)fluoranthene	1,52	NA	1000	52	12	1.03E+06	1.32E+09	43	365	NA NA
Chrysene	1.84	NA	1000	52	12	4.01E+07	1.32E+09	43	365	NA NA
Dibenz(a,h)anthracene	0.29	NA	1000	52	12	4.01E+07	1.32E+09	43	365	NA NA
Indeno(1,2,3-c,d)pyrene	1.15	NA	1000	52	12	3.66E+07	1.32E+09	43	365	NA
	 							Tota	l Pathway Risk	5E-09

Table 11 Risk Calculation for Dermal Contact with Gravel Fines Recreational Receptors Former Celotex Site - 2800 S. Sacramento Ave. Chicago, L.

Recreational Land Use - Gravel Fines

 $\label{eq:resolvent} \text{Risk}_{\text{soil-dem}} = \frac{C_{\text{soil}} \times \text{SF}_{\text{abs}} \times \text{CF}_2 \times \text{EF} \times \text{ED} \times \text{EV} \times \text{SA} \times \text{SSAF} \times \text{ABSd}}{\text{BW} \times \text{AT}_c \times \text{CF}_1}$ Equation: where: C_{soll} = Concentration in Soil (mg/kg) Calculated, 95% UCL See Table 2 SF_{aba} = Slope Factor, absorbed (SF_a ÷ ABS_{al}) See Table 3 ABS_o values are as presented in Exhibit 4-1 in RAGS Part E. CF₂ = Unit conversion Factor (kg/mg) 1E-06 EF = Exposure frequency (day/yr) 52 Conservative assumption (3 days/week during June, July, and August and 1 day/week during April, May, September, and October) ED = Exposure duration (yr) EV = Event Frequency (events/day) Recreational adolescent is assumed to range in age from 7 to 18. Therefore, total exposure duration is 12 years. 12 USEPA, 2004. RAGS, Part E, Exhibit 3-5. Default event frequency. SA = Skin Surface Area (cm²) 4,373 USEPA, 1997. Exposure Factors Handbook. Average surface area of head, hands, forearms, and lower legs of males and females aged 7-18 SSAF = Soil-to-skin Adherence Factor (mg/cm2-event) 0.07 USEPA, 2004 RAGS, Part E, Exhibit 3-5. Recommended soil-to-skin adherence factor for older children and adults, greater than 6 years of age. ABSd = Dermal Soil Absorption Factor (unitless) chemical-specific USEPA, 2004 RAGS, Part E. Exhibit 3-4 Recommended dermal absorption fraction from soil. USEPA, 1997. Exposure Factors Handbook, Table 7-3. Body weight is the average of males and females aged 7 to 18. BW≈ Body Weight (kg) 47 AT_c = Averaging Time (yr) TACO Default Value CF₁ = Unit conversion factor (days/year) 365

Chemical	Caon	SF ₀ ABS ₀	SF _{ate}	CF ₂	EF	ED	EV	SA	SSAF	ABSd	BW	ΑT _c	CF,	Risk _{epi-kerm}
	(mg/kg)	(mg/kg.day) (unitless)	(mg/kg day) ⁻¹	(kg/mg)	· (day/yr)	(ут)	(events/day)	(cm²)	(mg/cm²- event)	(unitless)	(kg)	(yr)	(day/yr)	ļ
	1		!											
Benz(a)anthracene	2.46	7:30E-01 1:00E+00	7.30E-01	1 0E-06	52	12	1	4,373	0.07	1.30E-01	47	70	365	3.71E-08
Benzo (a) pyrene	2.06	7.30E+00 1.00E+00	7 30E+00	1.0E-06	52	12	1	4,373	0 07	1.30E-01	47	70	365	3.11E-07
Benzo(b)fluoranthene	2.48	7.30E-01 1.00E+00	7.30E-01	1.0E-06	52	12	1	4,373	0.07	1.30E-01	47	70	365	3.74E-08
Benzo(k)fluoranthene	1.52	7/30E-02 1:00E+00	7.30E-02	1.0E-06	52	12	1	4,373	0 07	1.30E-01	47	70	365	2.30E-09
Chrysene	1.84	7.30E+03 1.00E+00	7.30E-03	1.0E-06	52	12	1	4,373	0.07	1.30E-01	47	70	365	2.78E-10
Dibenz(a,h)anthracene	0.29	7.30E+00 1.00E+00	7,30E+00	1.0E-06	52	12	1	4,373	0.07	1.30E-01	47	70	365	4.39E-08
Indeno (1,2,3-c,d)pyrene	1.15	730E-01 100E+00	7.30E-01	1.0E-06	52	12	1	4,373	0.07	1.30E-01	47	70	365	1.73E-0B
												Tota	l Pathway Risk	4E-07

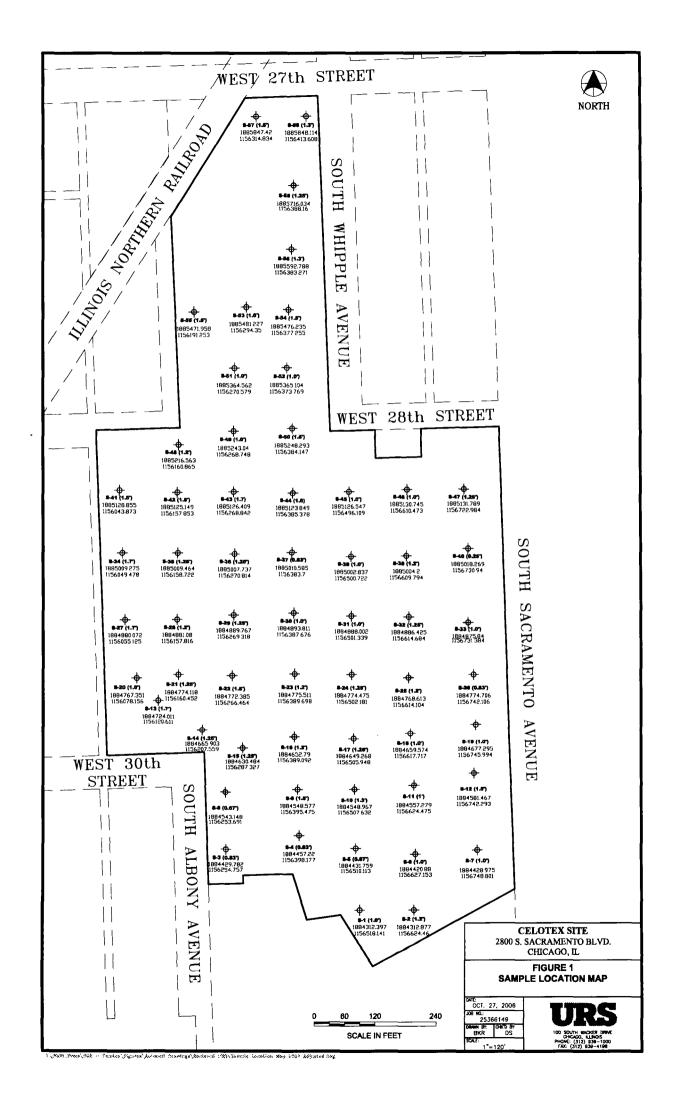
Table 12 Risk Characterization Summary Recreational Land Use Former Celotex Site - 2800 S. Sacramento Ave. Chicago, IL

Risk Summary - Gravel Fines

Exposure Receptor	Risk Estimate for	Risk Estimate for	Risk Estimate for	Total Risk
	Ingestion Pathway	Dermal Pathway ¹	Inhalation Pathway	Estimate
Residential	1E-06	4E-07	5E-09	2E-06

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Attachment A
Site Conceptual Exposure Model



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- 1.0 PHYSICAL AND CHEMICAL PROPERTIES OF COCS
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 - 3.1 Leaching to Groundwater
 - 3.2 Runoff to Surface Water Bodies
 - 3.3 Transport of Constituents Through Flooding
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- 4.0 POTENTIAL RECEPTORS
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 - 4.3 Potential Receptors for Groundwater-Related Exposure
- 5.0 POTENTIAL SOIL EXPOSURE ROUTES
- 6.0 ATTACHMENT B REFERENCES

A SCEM for the Site is established in this Attachment based on current understanding of site history, features, environmental settings, and future redevelopment plans. The SCEM reviews physical properties as well as potential fate and transport mechanisms of COCs identified in the TACO Tier 1 risk evaluation. The SCEM also characterizes potential sources, release mechanisms, migration pathways, potential receptors, exposure routes, and exposure pathway completeness.

1.0 PHYSICAL AND CHEMICAL PROPERTIES OF COCS

The following discussion reviews the physical properties, and potential fate and transport mechanisms of the COCs in the media where elevated concentrations were found. The COCs are constituents that were found in soil above the TACO Tier 1 soil ROs. The COCs identified at the Site are listed below.

Area/Media	COCs	Tier 1 ROs Exceeded	Approach to Address Tier 1 RO Exceedances	Addressed in
Surface Soil	Benzo(a)anthracene Benzo(a)pyrene Benzo(b)fluoranthene Dibenz(a,h)anthracene Indeno(1,2,3-cd)pyrene	Residential Ingestion	Tier 3 Formal Risk Assessment	Section 4.0 of this Report
Surface Soil	Benzo(a)anthracene Benzo(a)pyrene Benzo(b)fluoranthene Dibenz(a,h)anthracene Indeno(1,2,3-cd)pyrene Methylene chloride Dieldrin Chromium	Migration-to- Groundwater	Exposure Route Exclusion	Section 3.0 of this Report

Volatile Organic Chemicals

Methylene Chloride

Methylene chloride is used as solvent, chemical intermediate, grain fumigant, paint stripper and remover, metal degreaser, and refrigerant. If released to air, methylene chloride will exist solely as a vapor in the ambient atmosphere. Vapor-phase methylene chloride will be degraded in the atmosphere by reaction with photochemically-produced hydroxyl radicals. Methylene chloride will not be subject to direct photolysis. If released to soil, methylene chloride is expected to have very high mobility based upon its low organic carbon/water partition coefficient (Koc). Volatilization from moist soil surfaces is expected to be an important fate process based upon its Henry's Law constant. Methylene chloride may volatilize from dry soil surfaces based upon its vapor pressure. (HSDB, 2006).

Polycyclic Aromatic Hydrocarbons

PAHs are a class of organic compounds generally found in petroleum-derived products, asphalt, creosote oils, and coal products. In general, they have low vapor pressure, low solubility in water and high octanol water partition coefficients (Kow). They tend to be adsorbed to organic carbon in the soil, particularly the high molecular weight PAHs such as benz(a)anthracene,



benzo(b)fluoranthene, benzo(a)pyrene, dibenz(a,h)anthracene, and indeno(1,2,3-cd)pyrene which were potential COCs at the Site. Therefore, PAHs are not expected to be highly mobile in soil. PAHs also have slow biological degradation rates, which partly explains their persistence in soil and other media. Leaching to groundwater is not considered to be a significant pathway for PAHs, particularly in soils with higher organic carbon content. Most of the PAHs released to aquatic environments tend to remain near the sites of deposition (ATSDR, 1995).

Pesticide

Dieldrin

Dieldrin's former production and use as an insecticide resulted in its direct release to the environment. Dieldrin is also a degradation product of the insecticide aldrin, and the former use of aldrin has contributed to the occurrence of dieldrin in the environment. If released to air, dieldrin will exist in both the vapor and particulate phases in the ambient atmosphere. Vaporphase dieldrin will be degraded in the atmosphere by reaction with photochemically-produced hydroxyl radicals. Dieldrin also undergoes direct photolysis in the environment yielding photodieldrin as the primary degradation product. Particulate-phase dieldrin will be removed from the atmosphere by wet and dry deposition. If released to soil, dieldrin is expected to have low to no mobility based on its high organic carbon partition coefficient (Koc) value. Volatilization from moist soil surfaces is expected to be an important fate process based upon its Henry's Law constant; however adsorption may attenuate this process.

If released into water, dieldrin is expected to adsorb to suspended solids and sediment in water based upon its Koc data. Volatilization from water surfaces is expected to be an important fate process based upon this compound's Henry's Law constant. However, volatilization from water surfaces is expected to be attenuated by adsorption to suspended solids and sediment in the water column. (HSDB, 2006).

Metal

Chromium

Chromium is a metallic element with oxidation states ranging from chromium(-II) to chromium(+VI). The important valence states of chromium are trivalent (III) and hexavalent (VI). Chromium compounds are stable in the trivalent state and occur in nature in this state in ores, such as ferrochromite. The hexavalent is the second most stable state. However, hexavalent chromium rarely occurs naturally, but is produced from anthropogenic sources. Chromium is widely distributed in the earth's crust but is rare in unpolluted waters. The production and use of chromium compounds may result in their release to the environment through various waste streams. Chromium compounds are released into the atmosphere mainly by anthropogenetic stationary point sources, including industrial, commercial, and residential fuel combustion, via the combustion of natural gas, oil, and coal. If released to air, chromium compounds will exist solely in the particulate phase in the ambient atmosphere. Particulate-phase chromium compounds will be removed from the atmosphere by wet and dry deposition. If released to soil, the fate of chromium is greatly dependent upon the speciation of chromium, which is a function of the oxidation reduction potential (i.e., redox) and the pH of the soil. In most soils, chromium will be present predominantly in the trivalent state. This form has very low solubility and low reactivity resulting in low mobility in the environment. Under oxidizing conditions, hexavalent

Attachment A Site Conceptual Exposure Model

Receptor	Pathway			
Current Industrial/Commercial Workers	 Direct contact (i.e., incidental ingestion and dermal contact) with soils Inhalation of volatile organics and/or fugitive dusts 			
Future Recreational Users	 Direct contact (i.e., incidental ingestion and dermal contact) with soils Inhalation of volatile organics and/or fugitive dusts 			
Future Construction Workers	 Direct contact (i.e., incidental ingestion and dermal contact) with soils Inhalation of volatile organics and/or fugitive dusts 			

6.0 References

- Agency for Toxic Substances Disease Registry (ATSDR). 1995. *Toxicological Profile for Polycyclic Aromatic Hydrocarbons (PAHs)*. U.S. Department Of Health And Human Services Public Health Service. August.
- City of Chicago (1997). Memorandum of Understanding Between the City of Chicago and the Illinois Environmental Protection Agency. July.
- Illinois EPA (2001). *Tiered Approach to Corrective Action Objectives (TACO)*. Title 35, Section 742 of the Illinois Administrative Code.
- United States National Library of Medicine. 2006. *Hazardous Substance Data Bank*. http://toxnet.nlm.nih.gov/.

Table B-1 ProUCL Output

Chromium

Raw Statistics	_	Normal Distribution Test	
Number of Valid Samples	59	Lilliefors Test Statisitic	0.145002324
Number of Unique Samples	12	Lilliefors 5% Critical Value	0.115347375
Minimum	11	Data not normal at 5% significance level	
Maximum	25		
Mean	15.45763 _	95% UCL (Assuming Normal Distribut	<u>tion)</u>
Median	16	Student	16.03526094
Standard Deviation	2.654353		
Variance	7.045587	Gamma Distribution Test	
Coefficient of Variation	0.171718	A-D Test Statistic	0.828703063
Skewness	0.916414	A-D 5% Critical Value	0.748163556
		K-S Test Statistic	0.129629924
Gamma Statistics		K-S 5% Critical Value	0.115404203
k hat	36.33252	Data do not follow gamma distribution	
k star (bias corrected)	34.4964	at 5% significance level	
Theta hat	0.425449		
Theta star	0.448094 _	95% UCLs (Assuming Gamma Distribution	<u>on)</u>
nu hat	4287.237	Approximate Gamma UCL	16.03797878
nu star	4070.576	Adjusted Gamma UCL	16.05267238
Approx.Chi Square Value (.05)	3923.277		
Adjusted Level of Significance	0.045932	Lognormal Distribution Test	
Adjusted Chi Square Value	3919.686	Lilliefors Test Statisitic	0.139574601
		Lilliefors 5% Critical Value	0.115347375
Log-transformed Statistics	_	Data not lognormal at 5% significance level	
Minimum of log data	2.397895		
Maximum of log data	3.218876 _	95% UCLs (Assuming Lognormal Distri	<u>bution)</u>
Mean of log data	2.724278	95% H-UCL	16.04371921
Standard Deviation of log data	0.166501	95% Chebyshev (MVUE) UCL	16.92349134
Variance of log data	0.027723	97.5% Chebyshev (MVUE) UCL	17.55912019
		99% Chebyshev (MVUE) UCL	18.80768997
		0-0/ 1/	
	-	95% Non-parametric UCLs	
		CLT UCL	16.02603469
		Adj-CLT UCL (Adjusted for skewness)	16.07008804
		Mod-t U	16.04213237
		Jackknife UCL	16.03526094
		Standard Bootstrap UCL	16.04249712
DECOMMENDATION		Bootstrap-t UCL	16.05841749
RECOMMENDATION		Hall's Bootstrap UCL	16.09956764
Data are Non-parametric (0	(30.	Percentile Bootstrap UCL	16 01604015
Hoo Okindonio A LIOI		BCA Bootstrap UCL	16.01694915
Use Student's-t UCL		95% Chebyshev (Mean, Sd) UCL	16.96391991
or Modified-t UCL		97.5% Chebyshev (Mean, Sd) UCL	17.615694
		99% Chebyshev (Mean, Sd) UCL	18.895978

Table B-2 ProUCL Output

Benz(a)anthracene

Raw Statistics		Normal Distribution Test	
Number of Valid Samples	59 [–]	Lilliefors Test Statisitic	0.236523744
Number of Unique Samples	33	Lilliefors 5% Critical Value	0.115347375
Minimum	0.25	Data not normal at 5% significance level	
Maximum	9.9	•	
Mean	1.617966	95% UCL (Assuming Normal Distri	bution)
Median	1.1	Student's-t UCL	1.939955152
Standard Deviation	1.479609		
Variance	2.189244	Gamma Distribution Test	
Coefficient of Variation	0.914487	A-D Test Statistic	1.762496844
Skewness	3.703412	A-D 5% Critical Value	0.761380126
		K-S Test Statistic	0.174179842
Gamma Statistics		K-S 5% Critical Value	0.11703147
k hat	2.302614	Data do not follow gamma distribution	
k star (bias corrected)	2.196832	at 5% significance level	
Theta hat	0.702665		
Theta star	0.7365 _	95% UCLs (Assuming Gamma Distrib	
nu hat	271.7085	Approximate Gamma UCL	1.881332917
nu star	259.2261	Adjusted Gamma UCL	1.888438118
Approx.Chi Square Value (.05)	222.9372		
Adjusted Level of Significance	0.045932 _	Lognormal Distribution Test	
Adjusted Chi Square Value	222.0984	Lilliefors Test Statisitic	0.123543472
		Lilliefors 5% Critical Value	0.115347375
Log-transformed Statistics	-	Data not lognormal at 5% significance lev	/el
Minimum of log data	-1.386294		
Maximum of log data	2.292535 _	95% UCLs (Assuming Lognormal Di	
Mean of log data	0.248582	95% H-UCL	1.859766951
Standard Deviation of log data	0.64227	95% Chebyshev (MVUE) UCL	2.188875445
Variance of log data	0.412511	97.5% Chebyshev (MVUE) UCL	2.456944044
		99% Chebyshev (MVUE) UCL	2.983512897
		95% Non-parametric UCLs	
	-	CLT UCL	1.934812184
		Adj-CLT UCL (Adjusted for skewness)	2.034050071
		Mod-t UCL (Adjusted for skewness)	1.955434257
		Jackknife UCL	1.939955152
		Standard Bootstrap UCL	1.927760848
		Bootstrap-t UCL	2.153092142
RECOMMENDATION		Hall's Bootstrap UCL	3.402703754
Data are Non-parametric (0.0	05)	Percentile Bootstrap UCL	1.962033898
		BCA Bootstrap UCL	2.027118644
Use 95% Chebyshev (Mean,	Sd) UCL	95% Ch	2.457615315
		97.5% Chebyshev (Mean, Sd) UCL	2.8209322
		99% Chebyshev (Mean, Sd) UCL	3.53459787

Table B-4 ProUCL Output

Benzo(b)fluoranthene

Raw Statistics	_	Normal Distribution Test	
Number of Valid Samples	59	Lilliefors Test Statisitic	0.252019972
Number of Unique Samples	35	Lilliefors 5% Critical Value	0.115347375
Minimum	0.29	Data not normal at 5% significance level	
Maximum	8.7		
Mean	1.617797 _	95% UCL (Assuming Normal Distri	<u>bution)</u>
Median	1.2	Student's-t UCL	1.947780077
Standard Deviation	1.516346		
Variance	2.299304 _	Gamma Distribution Test	
Coefficient of Variation	0.937291	A-D Test Statistic	2.197791626
Skewness	3.017568	A-D 5% Critical Value	0.762258828
		K-S Test Statistic	0.190232599
Gamma Statistics		K-S 5% Critical Value	0.117128317
k hat	2.125563	Data do not follow gamma distribution	
k star (bias corrected)	2.028783	at 5% significance level	
Theta hat	0.761114		
Theta star	0.797422 _	95% UCLs (Assuming Gamma Distrib	
nu hat	250.8165	Approximate Gamma UCL	1.89321232
nu star	239.3964	Adjusted Gamma UCL	1.900669022
Approx.Chi Square Value (.05)	204.5702		
Adjusted Level of Significance	0.045932 _	Lognormal Distribution Test	
Adjusted Chi Square Value	203.7676	Lilliefors Test Statisitic	0.140173255
		Lilliefors 5% Critical Value	0.115347375
Log-transformed Statistics	_	Data not lognormal at 5% significance lev	el
Minimum of log data	-1.237874		
Maximum of log data	2.163323 _	95% UCLs (Assuming Lognormal Di	
Mean of log data	0.227761	95% H-UCL	1.856870459
Standard Deviation of log data	0.661991	95% Chebyshev (MVUE) UCL	2.192526505
Variance of log data	0.438232	97.5% Chebyshev (MVUE) UCL	2.46776414
		99% Chebyshev (MVUE) UCL	3.008415176
		OF9/ Non novement HOLO	
	-	95% Non-parametric UCLs CLT UCL	1.942509418
			2.025376813
		Adj-CLT UCL (Adjusted for skewness) Mod-t UCL (Adjusted for skewness)	1.960705716
		Jackknife UCL	1.947780077
		Standard Bootstrap UCL	1.946036835
		Bootstrap-t UCL	2.100829402
RECOMMENDATION		Hall's Bootstrap UCL	2.12267588
Data are Non-parametric (0.0	05)	Percentile Bootstrap UCL	1.951355932
zata are rion paramonio (o.	,	BCA Bootstrap UCL	2.039152542
Use 95% Chebyshev (Mean,	Sd) UCI	95% Ch	2.478292821
333 33.3 Shobyonov (mount,	o, ool	97.5% Chebyshev (Mean, Sd) UCL	2.850630218
		99% Chebyshev (Mean, Sd) UCL	3.582014937
		ob to Onobyonov (Moan, Out OOL	3.30 <u>2</u> 01 7 337

Table B-5 ProUCL Output

Dibenz(a,h)anthracene

Raw Statistics	_	Normal Distribution Test	
Number of Valid Samples	59	Lilliefors Test Statisitic	0.234620247
Number of Unique Samples	30	Lilliefors 5% Critical Value	0.115347375
Minimum	0.073	Data not normal at 5% significance level	
Maximum	1		
Mean	0.25439 _	95% UCL (Assuming Normal Distril	oution)
Median	0.19	Student's-t UCL	0.298096159
Standard Deviation	0.20084		
Variance	0.040337 _	Gamma Distribution Test	
Coefficient of Variation	0.789497	A-D Test Statistic	1.697930781
Skewness	2.389727	A-D 5% Critical Value	0.760104166
		K-S Test Statistic	0.149149784
Gamma Statistics		K-S 5% Critical Value	0.116890839
k hat	2.559709	Data do not follow gamma distribution	
k star (bias corrected)	2.440853	at 5% significance level	
Theta hat	0.099382		
Theta star	0.104222 _	95% UCLs (Assuming Gamma Distrib	
nu hat	302.0456	Approximate Gamma UCL	0.293425239
nu star	288.0207	Adjusted Gamma UCL	0.294473649
Approx.Chi Square Value (.05)	249.7043		
Adjusted Level of Significance	0.045932 _	Lognormal Distribution Test	
Adjusted Chi Square Value	248.8153	Lilliefors Test Statisitic	0.099264777
		Lilliefors 5% Critical Value	0.115347375
Log-transformed Statistics	_	Data are lognormal at 5% significance lev	el
Minimum of log data	-2.617296		
Maximum of log data	0 _	95% UCLs (Assuming Lognormal Dis	
Mean of log data	-1.576759	95% H-	0.291017969
Standard Deviation of log data	0.611055	95% Chebyshev (MVUE) UCL	0.34069401
Variance of log data	0.373389	97.5% Chebyshev (MVUE) UCL	0.380744741
		99% Chebyshev (MVUE) UCL	0.459416649
		95% Non-parametric UCLs	
	_	CLT UCL	0.29739806
		Adj-CLT UCL (Adjusted for skewness)	0.306090205
		Mod-t UCL (Adjusted for skewness)	0.299451958
		Jackknife UCL	0.298096159
		Standard Bootstrap UCL	0.297334655
		Bootstrap-t UCL	0.314077099
RECOMMENDATION		Hall's Bootstrap UCL	0.306282126
Data are lognormal (0.05)		Percentile Bootstrap UCL	0.296644068
		BCA Bootstrap UCL	0.305220339
Use H-UCL		95% Chebyshev (Mean, Sd) UCL	0.368362602
		97.5% Chebyshev (Mean, Sd) UCL	0.417678712
		99% Chebyshev (Mean, Sd) UCL	0.514550659
Data are lognormal (0.05)		Hall's Bootstrap UCL Percentile Bootstrap UCL BCA Bootstrap UCL 95% Chebyshev (Mean, Sd) UCL	0.306282126 0.296644068 0.305220339 0.368362602
		99% Chebyshev (Mean, Sd) UCL	

Table B-6 ProUCL Output

Indeno(1,2,3-cd)pyrene

Raw Statistics		Normal Distribution Test	
Number of Valid Samples	59	Lilliefors Test Statisitic	0.272585607
Number of Unique Samples	43	Lilliefors 5% Critical Value	0.115347375
Minimum	0.14	Data not normal at 5% significance level	
Maximum	3.7		
Mean	0.720508	95% UCL (Assuming Normal Distri	bution)
Median	0.49	Student's-t UCL	0.8838649
Standard Deviation	0.750658		
Variance	0.563488	Gamma Distribution Test	
Coefficient of Variation	1.041845	A-D Test Statistic	3.564424148
Skewness	2.860216	A-D 5% Critical Value	0.764426835
		K-S Test Statistic	0.211900434
Gamma Statistics		K-S 5% Critical Value	0.11737918
k hat	1.867985	Data do not follow gamma distribution	
k star (bias corrected)	1.784302	at 5% significance level	
Theta hat	0.385714		
Theta star	0.403804 _	95% UCLs (Assuming Gamma Distrib	ution)_
nu hat	220.4223	Approximate Gamma UCL	0.852435097
nu star	210.5477	Adjusted Gamma UCL	0.856028648
Approx.Chi Square Value (.05)	177.9624		
Adjusted Level of Significance	0.045932 _	Lognormal Distribution Test	<u> </u>
Adjusted Chi Square Value	177.2153	Lilliefors Test Statisitic	0.164557123
		Lilliefors 5% Critical Value	0.115347375
Log-transformed Statistics		Data not lognormal at 5% significance lev	rel .
Minimum of log data	-1.966113		
Maximum of log data	1.308333 _	95% UCLs (Assuming Lognormal Di	
Mean of log data	-0.618738	95% H-UCL	0.814925294
Standard Deviation of log data	0.68485	95% Chebyshev (MVUE) UCL	0.965697967
Variance of log data	0.46902	97.5% Chebyshev (MVUE) UCL	1.090323434
		99% Chebyshev (MVUE) UCL	1.33512603
		95% Non-parametric UCLs	
	_	CLT UCL	0.88125569
		Adj-CLT UCL (Adjusted for skewness)	0.920139559
		Mod-t UCL (Adjusted for skewness)	0.889929998
		Jackknife UCL	0.8838649
		Standard Bootstrap UCL	0.875544295
		Bootstrap-t UCL	0.959905367
RECOMMENDATION		Hall's Bootstrap UCL	0.927953155
Data are Non-parametric (0.	05)	Percentile Bootstrap UCL	0.882033898
	•	BCA Bootstrap UCL	0.941016949
Use 95% Chebyshev (Mean,	Sd) UCL	95% Ch	1.146492195
•		97.5% Chebyshev (Mean, Sd) UCL	1.330815692
		99% Chebyshev (Mean, Sd) UCL	1.692883508
		· · ·	

Table B-7 ProUCL Output

Benzo(k)fluoranthene

Raw Statistics		Normal Distribution Test	
Number of Valid Samples	59	Lilliefors Test Statisitic	0.292122763
Number of Unique Samples	45	Lilliefors 5% Critical Value	0.115347375
Minimum	0.14	Data not normal at 5% significance level	
Maximum	6.4	· ·	
Mean	0.945424	95% UCL (Assuming Normal Distri	bution)
Median	0.69	Student's-t UCL	1.166348863
Standard Deviation	1.015199		
Variance	1.030629	Gamma Distribution Test	
Coefficient of Variation	1.073803	A-D Test Statistic	2.896198052
Skewness	3.756763	A-D 5% Critical Value	0.763279775
		K-S Test Statistic	0.193607285
Gamma Statistics		K-S 5% Critical Value	0.117243909
k hat	1.966008	Data do not follow gamma distribution	
k star (bias corrected)	1.877341	at 5% significance level	
Theta hat	0.480885		
Theta star	0.503597 _	95% UCLs (Assuming Gamma Distrib	
nu hat	231.9889	Approximate Gamma UCL	1.113601511
nu star	221.5262	Adjusted Gamma UCL	1.118171295
Approx.Chi Square Value (.05)	188.071		
Adjusted Level of Significance	0.045932 _	Lognormal Distribution Test	
Adjusted Chi Square Value	187.3024	Lilliefors Test Statisitic	0.128958034
		Lilliefors 5% Critical Value	0.115347375
Log-transformed Statistics	_	Data not lognormal at 5% significance lev	/el
Minimum of log data	-1.966113		
Maximum of log data	1.856298 _	95% UCLs (Assuming Lognormal Di	
Mean of log data	-0.331502	95% H-UCL	1.070610664
Standard Deviation of log data	0.670636	95% Chebyshev (MVUE) UCL	1.265881281
Variance of log data	0.449753	97.5% Chebyshev (MVUE) UCL	1.426482834
		99% Chebyshev (MVUE) UCL	1.741953486
		95% Non-parametric UCLs	
		CLT UCL	1.162820138
		Adj-CLT UCL (Adjusted for skewness)	1.23189076
		Mod-t UCL (Adjusted for skewness)	1.177122485
		Jackknife UCL	1.166348863
		Standard Bootstrap UCL	1.156921324
		Bootstrap-t UCL	1.318180542
RECOMMENDATION		Hall's Bootstrap UCL	1.417142651
Data are Non-parametric (0.	05)	Percentile Bootstrap UCL	1.181694915
		BCA Bootstrap UCL	1.248474576
Use 95% Chebyshev (Mean,	Sd) UCL	95% Ch	1.521529081
		97.5% Chebyshev (Mean, Sd) UCL	1.770810327
		99% Chebyshev (Mean, Sd) UCL	2.260475065

Table B-8 ProUCL Output

Chrysene

Raw Statistics_	_	Normal Distribution Test	
Number of Valid Samples	59	Lilliefors Test Statisitic	0.221902895
Number of Unique Samples	36	Lilliefors 5% Critical Value	0.115347375
Minimum	0.25	Data not normal at 5% significance level	
Maximum	8.6		
Mean	1.604746 _	95% UCL (Assuming Normal Distri	<u>bution)</u>
Median	1.2	Student's-t UCL	1.895553221
Standard Deviation	1.336323		
Variance	1.78576 _	Gamma Distribution Test	
Coefficient of Variation	0.832732	A-D Test Statistic	1.445126365
Skewness	3.257054	A-D 5% Critical Value	0.760280014
		K-S Test Statistic	0.160390013
Gamma Statistics		K-S 5% Critical Value	0.116910221
k hat	2.524277	Data do not follow gamma distribution	
k star (bias corrected)	2.407223	at 5% significance level	
Theta hat	0.635725		
Theta star	0.666638 _	95% UCLs (Assuming Gamma Distrib	
nu hat	297.8647	Approximate Gamma UCL	1.852906175
nu star	284.0523	Adjusted Gamma UCL	1.859575079
Approx.Chi Square Value (.05)	246.0091		
Adjusted Level of Significance	0.045932 _	Lognormal Distribution Test	
Adjusted Chi Square Value	245.1269	Lilliefors Test Statisitic	0.113185489
		Lilliefors 5% Critical Value	0.115347375
Log-transformed Statistics		Data are lognormal at 5% significance lev	el
Minimum of log data	-1.386294		
Maximum of log data	2.151762 _	95% UCLs (Assuming Lognormal Dis	
Mean of log data	0.262003	95% H-I	1.844790081
Standard Deviation of log data	0.619629	95% Chebyshev (MVUE) UCL	2.162919506
Variance of log data	0.38394	97.5% Chebyshev (MVUE) UCL	2.420118427
		99% Chebyshev (MVUE) UCL	2.925335903
	_	95% Non-parametric UCLs	
	_	CLT UCL	1.8909083
		Adj-CLT UCL (Adjusted for skewness)	1.969733473
		Mod-t UCL (Adjusted for skewness)	1.907848355
		Jackknife UCL	1.895553221
		Standard Bootstrap UCL	1.887560541
		Bootstrap-t UCL	2.034269481
RECOMMENDATION		Hall's Bootstrap UCL	2.375897253
Data are lognormal (0.05)		Percentile Bootstrap UCL	1.906779661
		BCA Bootstrap UCL	1.961864407
Use H-UCL		95% Chebyshev (Mean, Sd) UCL	2.363082894
		97.5% Chebyshev (Mean, Sd) UCL	2.691215977
		99% Chebyshev (Mean, Sd) UCL	3.335769879